

A review on health and environmental aspects of current manure management practices in pig production systems in Uganda

Ibrahim Wanyama and Sonja Leitner
International Livestock Research Institute

December 2019



©2019 International Livestock Research Institute (ILRI)

ILRI thanks all donors and organizations which globally support its work through their contributions to the [CGIAR Trust Fund](#)



This publication is copyrighted by the International Livestock Research Institute (ILRI). It is licensed for use under the Creative Commons Attribution 4.0 International Licence. To view this licence, visit <https://creativecommons.org/licenses/by/4.0>.

Unless otherwise noted, you are free to share (copy and redistribute the material in any medium or format), adapt (remix, transform, and build upon the material) for any purpose, even commercially, under the following conditions:

 **ATTRIBUTION.** The work must be attributed, but not in any way that suggests endorsement by ILRI or the author(s).

NOTICE:

For any reuse or distribution, the licence terms of this work must be made clear to others.

Any of the above conditions can be waived if permission is obtained from the copyright holder.

Nothing in this licence impairs or restricts the author's moral rights.

Fair dealing and other rights are in no way affected by the above.

The parts used must not misrepresent the meaning of the publication.

ILRI would appreciate being sent a copy of any materials in which text, photos etc. have been used.

Editing, design and layout—ILRI Editorial and Publishing Services, Addis Ababa, Ethiopia.

Cover photo—ILRI/Karen Marshall

Citation: Wanyama, I. and Leitner, S. 2019. *A review on health and environmental aspects of current manure management practices in pig production systems in Uganda*. Nairobi, Kenya: ILRI.

Patron: Professor Peter C Doherty AC, FAA, FRS

Animal scientist, Nobel Prize Laureate for Physiology or Medicine—1996

Box 30709, Nairobi 00100 Kenya

Phone +254 20 422 3000

Fax +254 20 422 3001

Email ilri-kenya@cgiar.org

ilri.org

better lives through livestock

ILRI is a CGIAR research centre

Box 5689, Addis Ababa, Ethiopia

Phone +251 11 617 2000

Fax +251 11 667 6923

Email ilri-ethiopia@cgiar.org

ILRI has offices in East Africa • South Asia • Southeast and East Asia • Southern Africa • West Africa

Contents

Summary	ii
Background	1
Greenhouse gas emissions from manure	2
Pig production systems in Uganda	3
Pig manure characteristics	3
Pig manure and health issues	4
Manure handling, storage and treatment options to reduce GHG emissions and improve health and safety	6
Manure storage	6
Manure composting	7
Biogas production	7
Vermicomposting	8
Conclusion	9
References	10

Summary

Pig production in Uganda has increased over the past two decades and is projected to grow further due to the rising demand for pork. Pig production creates substantial amounts of manure and, if not handled well, can be a source of greenhouse gas (GHG) and ammonia (NH₃) emissions, as well as nutrient leaching, resulting in acidification of rainwater and eutrophication of water bodies. Furthermore, pig manure can be both host and vector for diseases. To manage pig manure in a way that is both climate smart (i.e. helps to mitigate climate change) and sanitary, it is essential to have a detailed understanding of manure management systems that comprises animal feeds (diets), production systems, storage and handling, treatment, manure use and disposal.

This report presents the results of a literature review conducted to assess current pig production systems and associated manure management in Uganda, while also referring to literature from elsewhere where no information for Uganda was available. We specifically focused on the environmental (GHG emissions) and health (disease transmission) aspects of pig manure in Uganda. We found that pig production systems in Uganda can be categorized as either extensive, semi-intensive (tethering) and intensive. However, literature on pig manure management in Uganda is very scarce as most of the studies done in the country only characterized production systems but did not provide any information on how manure is managed. Across the country, pig manure is perceived as waste and its management and usefulness in improving soil fertility is not included in the Uganda fertilizer policy. However, given the growth of pig production, there is need to carry out studies to establish how manure is managed in different production systems, which can eventually guide the development of smart management practices that are environmentally friendly, assure high fertilizer quality and reduce animal and human health risks. Such studies could be instrumental in guiding policy advocacy and development in Uganda.

Background

There has been remarkable increase in global pig production in the past decade, especially in developing countries, to meet the rising demand from population growth and urbanization, as well as change in dietary preference towards more animal-based protein (Thornton 2010). Furthermore, pigs provide quicker returns than cattle because of their shorter lifecycle (Ndyomugenyi and Kyasimire 2015) providing income to meet the immediate needs of farmers. Pig production also requires relatively little space compared to other livestock production systems, e.g. extensive cattle production, which makes it suitable for peri-urban settings (Thys et al. 2016).

The Uganda Bureau of Statistics noted a rapid increase in the number of pigs from 0.19 million in 1980 to 3.2 million in 2008 (MAAIF 2008). Eighteen percent of all households (1.1 million) own at least one pig, with the greatest pig population being located in the central districts (Wakiso, Masaka and Mukono), while the rest of the country is dominated by smallholder pig farmers (Ikwap et al. 2014; Anderson et al. 2016). However, the fast growth of pig production in Uganda poses several environmental, health and safety challenges. For example, waste from pig production such as manure produces a foul smell which can create conflict in urban and peri-urban pig production (Alvåsen 2007; Nantima et al. 2015). Furthermore, pig manure emits ammonia (NH₃) that causes acid rain (Kruse and Bell 1987; Burns et al. 2016), and phosphorus leaching from manure results in eutrophication of water bodies (Daniel et al. 1998). Depending on the management system, pig manure can also be a substantial source of greenhouse gas (GHG) emissions (Herrero et al. 2016; Sakadevan and Nguyen 2017). Several studies have also observed that pig manure hosts diseases that can be transmitted to humans and other livestock (e.g. Bicudo and Goyal 2003; Ström et al. 2018; Venglovsky et al. 2018).

The present review aims to characterize current pig production systems in central Uganda with a focus on existing manure management and its environmental, health and safety issues. Targeted recommendations to improve manure management to reduce environmental and health burdens (“smart manure management”) will also be provided.

Greenhouse gas emissions from manure

Globally, the agriculture, forestry and other land uses (AFOLU) sector is the second largest contributor of GHG emissions with 23% (12 Gt CO₂eq yr⁻¹) of the total annual net anthropogenic GHG emissions (52 Gt CO₂eq yr⁻¹) for the period of 2007–2016 (IPCC 2019a, p.10). Concerning global livestock GHG emissions, cattle emissions dominate by far (54% from beef cattle and 17% from dairy cattle), followed by emissions from sheep (9%), buffalo (7%), pigs (5%) and goats (4%) (Caro et al. 2014). Of total GHG emissions from pig production globally, pig manure emissions represent 35.3% and include emissions from manure storage and processing (27.4%), and from applied and deposited manure (Macleod et al. 2013). With regards to climate change mitigation, the livestock sector is estimated to have a technical mitigation potential of up to 50% of the GHG emissions of the AFOLU sector, of which reduction of GHG emissions from manure is estimated at 0.01–0.08 GtCO₂eq yr⁻¹, mainly through reduction of nitrous oxide (N₂O) emissions and nitrogen (N) losses from manure (Herrero et al. 2016).

However, these estimates have large uncertainties due to lack of accurate localized activity data (e.g. livestock numbers, feed availability and quality) from sub-Saharan Africa (SSA). Estimates from SSA are often based on default emission factors which have been largely generated in developed countries where conditions are quite different from SSA (e.g. different breeds, feeds and climatic conditions). In order to improve the reporting of national GHG inventories to fulfill the countries' nationally determined contributions (NDC) as agreed upon following the Paris Climate Agreement, a number of empirical studies have been carried out in East Africa to improve these estimates in Uganda (e.g. Kiggundu et al. 2019) and Kenya (e.g. Pelster et al. 2016; Goopy et al. 2018; Ndung'u et al. 2018; Zhu et al. 2018), but their focus has been mainly on cattle. Research that focuses on GHG emissions from pig production systems based in East Africa on in situ data remains scarce.

Pig production systems in Uganda

The primary factors that influence GHG emissions from manure include temperature, oxygen (O₂) availability (aeration), moisture and sources of nutrients. In addition, emissions are affected by manure type (livestock type/breed), feed type, manure texture (e.g. liquid slurry, solid, litter-based farmyard manure), manure storage type (e.g. solid heap, anaerobic lagoon, pit under confinement), and manure application to soils (e.g. top-dressed, injected, incorporated). These factors are largely influenced by the livestock production system (Robinson et al. 2011; Petersen et al. 2013). In Uganda, pig production systems can be separated into (i) extensive, (ii) semi-intensive (tethering) and (iii) intensive (FAO 2012; Kimbi et al. 2015; Nantima et al. 2015; Ndyomugenyi and Kyasimire 2015; Pfeifer et al. 2015).

i. Extensive pig production system

In this system, animals are left to scavenge during the day and are confined either in pens, in other simple constructions, or tethered during the night. This system is mostly practiced in low input systems in rural areas where scavenging space is available (Thornton 2010; Dione et al. 2014), and is commonly applied during the dry season when feeds are scarce and expensive to buy (Dione et al. 2014). In this system, manure is deposited randomly on pastures while pigs scavenge during the day, making manure collection difficult (Pfeifer et al. 2015). Manure that is deposited in animal pens or fenced confinements overnight can be collected.

ii. Semi-intensive pig production system

This system is also referred to as tethering system, where pigs are usually tied with a rope under a tree or in a garden next to the home and left to scavenge within their movement radius. They are often supplemented with water, home refuse and crop residues. Once pastures are exhausted, pigs are moved to a different spot. This system is the most common system in rural areas especially in upcountry districts (Tororo, Hoima, Kayunga, Kamuli) that produce pigs for rural and urban markets (Dione et al. 2014; Nalubwama et al. 2014; Nantima et al. 2015; Pfeifer et al. 2015; Roesel et al. 2017). In this system, manure is produced in a restricted area and can be managed more easily as compared to a free-range system. A focus group discussion in Hoima district revealed that when farmers periodically move animals once the pasture is exhausted, they spread and incorporate the manure into the soil as fertilizer (Pfeifer et al. 2015).

iii. Intensive pig production system

This is a high-investment system where animals are totally confined and provided with shelter, food and water. This system is used where there is not enough land to practice free-range or tethering and is common in urban areas and among rural farmers who produce intensively for the urban market (Dione et al. 2014; Pfeifer et al. 2015). Herd sizes in this system are often large (50–300 animals) compared to free-range or tethering systems (one to five animals) (Okello et al. 2015; Nabiky and Kugonza 2016). Many farmers in intensive production systems keep improved pig breeds, which are more susceptible to tropical diseases (Dione et al. 2014; Pfeifer et al. 2015). Pigs in intensive systems are usually fed on nutrient- and energy-rich commercial feeds such as concentrates, maize and rice brans, and waste. In these systems, animals are housed fulltime, in most cases in barns with raised or concrete floors that can facilitate manure collection, which is primarily done for hygiene purposes (Dione et al. 2014). Key stakeholders in Hoima district reported that in this system, manure is mostly washed into pits or stacked into piles and is often not used as fertilizer (Pfeifer et al. 2015). Another study in Kampala found that 48% of farmers apply manure from their pig units in their gardens, 30 % heap the manure and dispose of it in garbage dumping areas or other unspecified areas after a certain storage period (ranging from one week to 3 months), and 10 % of the pig farmers dispose the manure daily in dumping sites or in other unspecified areas (Alvåsen 2007).

Pig manure characteristics

Pig manure consists of faeces and urine and is rich in carbon (C), nitrogen (N, primarily urea from urine and organic N from faeces), phosphorus (P), potassium (K), calcium (Ca), magnesium (Mg), sodium (Na), and micro-elements such as iron (Fe), manganese (Mn), zinc (Zn), cobalt (Co) and copper (Cu) (Okoli et al. 2019). Pig manure generally has higher N concentrations compared to manure from ruminants such as cattle that feed on roughages (Table I). Consequently, the carbon: nitrogen ratio (C:N ratio) of pig manure is lower than cattle manure, indicating a higher fertilizer value and better degradability. For this reason, pig manure has been reported to yield more biogas compared to cattle manure (Fantozzi and Buratti 2009; Li et al. 2020).

Source	Nigeria (Okoli et al. 2019)	Uganda (Alvåsen 2007)		Zimbabwe (Nyamangara et al. 2010)	
Manure type	Fresh pig manure	Fresh pig manure	Fresh cattle manure	Fresh pig manure	Fresh cattle manure
Nitrogen (% DM)	2.25	3.5	1	3.1	0.8
C:N ratio	7.8	11.8	31.8	4.3	14.3

The concentration of nutrients in faecal matter and urine depends largely on the animal's diet. For instance, faecal and urinary N excretion from pigs has been found to increase with increasing dietary crude protein (CP) intake (Portejoie et al. 2004; Liu et al. 2018). Furthermore, Liu et al. (2018) observed a linear relationship between faecal P excretion and dietary P intake as long as the Ca:P ratio remained constant. High levels of Ca in diets reduce P digestibility through formation of an insoluble Ca-P complex, and thus most of the dietary P intake is excreted (Selle et al. 2009). The amount of N and P excreted in faecal matter and urine is further determined by several other factors such as animal growth stage, feed formulation, pig species and complex interactions between these factors. Dourmad et al. (1999) observed that mean N retention in pigs differed among growing stages with 23% dietary N being retained by sows, 34% by growing pigs and 48% by weaners. Furthermore, partitioning of N between faecal and urinary excreta is also influenced by diet composition (Nguyen et al. 2019) with diets high in dietary fibre (cellulose and hemicelluloses) leading to more allocation of N to faeces compared to urine (Prapasongsa et al. 2009). Since dietary fibre is the main factor for N partitioning between faeces and urine (Jørgensen et al. 2013), dietary manipulation towards higher fibre intake can be used to shift the allocation from urinary N (mostly urea) towards faecal N (mostly organic N), which can help to reduce NH₃ emissions from pig manure (Galassi et al. 2010). Additionally, diet composition also affects the pH of faeces, urine and their mixture (Canh et al. 1997). Both N partitioning and pH modification affect NH₃ and GHG emissions from manure. As mentioned before, NH₃ emissions can cause acid rain and lead to soil acidification (Guthrie et al. 2018). Furthermore, NH₃ reacts with acids in the atmosphere, which leads to formation of ammonium aerosols—fine particulate matter (aerodynamic diameter <2.5 µm) that poses a health hazard (Aneja et al. 2008). Given that pig production in Uganda is growing rapidly, especially in peri-urban and urban areas where commercial nutrient-rich feeds are used (MAAIF 2008), this could exacerbate environmental pollution unless mitigation measures are put in place.

Pig manure and health issues

Pig manure contains microorganisms, protozoa and viruses that may pose a risk to human and animal health. A recent review by Alegbeleye et al. (2018) highlighted public health diseases that can be transmitted through raw or poorly treated animal manure, including *Escherichia coli* O157:H7, *Salmonella* spp., *Listeria monocytogenes*, *Campylobacter* spp., porcine enteroviruses, bovine coronavirus, bovine virus, diarrhea, *Cryptosporidium parvum* and *Giardia*. In addition, contaminated soils can act as a habitat for pathogens with some persisting for months or even years in the soil (Alegbeleye et al. 2018). In Uganda, it is common practice to spread untreated manure on the soil, which represents high risk of pathogen distribution in the environment. Highly contagious and pathogenic viral diseases such as foot and mouth disease, classical swine fever and Aujeszky's disease may spread through animal effluents into waterways and

when one farm is infected, other downstream farms may be exposed to considerable risk of infection (Johansson et al. 2005). Pathogens in untreated manure can also be transmitted to other livestock, humans or food by vectors such as flies, rodents and birds (von Holy et al. 2006) This is a common problem in systems where manure is not collected routinely. Furthermore, use of antibiotics to control bacterial infections is increasing in Ugandan pig production which poses the risk of development of antibiotic-resistant bacteria in pigs, which may further lead to emergence of antibiotic resistance in human and other livestock pathogens (Kang et al. 2018; Xie et al. 2018). It is estimated that 70–90% of antibiotics administered to pigs are discharged via faeces and urine in unchanged or active metabolite form (Dolliver et al. 2008). However, treatment of resistant strains is expensive, often more toxic for the host, and in general difficult to treat if suitable drugs are not available.

Manure handling, storage and treatment options to reduce GHG emissions and improve health and safety

GHGs from manure are produced via different microbial pathways. These processes are influenced by manure characteristics (e.g. water content, O_2 level, nutrient concentration) and environmental parameters (e.g. temperature, rainfall). N_2O is mainly produced through nitrification and denitrification processes, although other pathways such as dissimilatory nitrate reduction to ammonium (DNRA) have been reported (Butterbach-Bahl et al. 2013). Nitrification involves the microbially mediated transformation of ammonium (NH_4^+) to nitrate (NO_3^-) by ammonium- and nitrite-oxidizing bacteria and archaea under aerobic conditions (i.e. presence of O_2). On the other hand, denitrification is a (facultatively) anaerobic process (i.e. requires low levels or absence of O_2) and involves reduction of NO_3^- and nitrite (NO_2^-) to N_2O and N_2 gas. In addition to low O_2 levels, denitrifier microorganisms are heterotrophic and require C as a source of energy. Nitrification and denitrification are closely linked, and N_2O emissions have been reported to be highest under intermediate water content, when both aerobic and anaerobic microsites exist in proximity (Butterbach-Bahl et al. 2013). Furthermore, N_2O formation responds to temperature, with the optimum temperature being between 20 and 35°C, which leads to high emissions in tropical and subtropical regions.

Contrary to N_2O formation, CH_4 production (methanogenesis) is a strictly anaerobic process that requires absence of O_2 and the presence of C substrates from faecal matter or bedding material that can be reduced to CH_4 (Serrano-Silva et al. 2014). Similar to N_2O formation, CH_4 production is controlled by temperature, moisture and availability of organic substrates. All these factors—moisture, temperature and C availability—are in turn affected by the type of manure management (i.e. handling, treatment and storage conditions) (Petersen et al. 2013). Pig manure is handled as solid, semi-solid/slurry or liquid depending on the type of housing and production system. Generally, solid manure (including deposited manure on pastures) produces less CH_4 and more N_2O , while manure handled in slurry or liquid form produces more CH_4 but less N_2O (Chadwick et al. 2011).

Manure storage

Manure storage GHG emissions represent 10–25% of total emissions from the livestock sector (Gerber et al. 2013; Serrano-Silva et al. 2014). In pig production systems, manure storage is the greatest source of GHG emissions (69%) since pigs are monogastric animals and do not produce enteric CH_4 like ruminants (cattle, sheep and goats). Emissions of GHGs and NH_3 during storage are influenced by water content of manure (solid, slurry, or liquid), environmental factors (e.g. temperature, humidity and wind), and the storage duration. In smallholder and extensive systems in SSA, manure is mainly stored in solid form (Petersen et al. 2013). Manure that is dropped when animals are tethered or confined is collected in heaps and applied to fields after varying durations of storage times, or disposed of as waste (Ndambi et al. 2019). Manure from free-roaming animals is not usually collected; it's left where it is deposited. However, manure that is left in open fields without management is prone to nutrient loss through leaching and volatilization, thereby compromising fertilizer value and promoting GHG emissions. Furthermore, unmanaged manure is a liability as it can promote diseases.

A study on urban and peri-urban livestock farming from Tanzania observed that most of the respondents stored manure in heaps before disposal, while only a few farmers directly spread the manure in the surrounding environment (Lupindu et al. 2012). Similar observations were made in Hoima, Uganda (Alvåsen 2007). Heaped solid manure primarily emits N_2O given the conducive environment of partial aerobic and anaerobic state that promotes nitrification and denitrification, but if the heap is big enough to maintain a moist, anaerobic core, it can also promote

CH₄ emissions (Chadwick et al. 2011). Therefore, management practices that increase aeration, such as regular turning to promote O₂ supply, can help reduce N₂O and CH₄ emissions from manure (Petersen et al. 2013).

In the intensive pig production systems of central Uganda, some farmers house their animals on concrete floors which are flushed with water for cleaning. Pig manure is collected after flushing and stored in liquid/slurry form (Alvåsen 2007; Pfeifer et al. 2015). Manure stored as liquid/slurry represents an anaerobic environment, which promotes CH₄ emissions. Since temperature is a driving factor for CH₄ production (Sommer et al. 2007; Elsgaard et al. 2016; Moset et al. 2019), CH₄ production from slurry lagoons is promoted under subtropical conditions. While CH₄ emissions can be decreased via a reduction of storage temperature (Borhan et al. 2012), temperature manipulation to mitigate CH₄ emissions from liquid/slurry manure in SSA might be challenging. Other than temperature, factors such as microbial community composition, manure organic matter quality content and pH also affect CH₄ production from liquid/slurry manure (Elsgaard et al. 2016). However, these factors are difficult to manipulate in smallholder settings and reducing CH₄ emissions from slurry/liquid manure storage in Uganda might currently be challenging.

Manure composting

Composting is a widely recognized biological process of managing organic waste. Composting is the microbially mediated bio oxidation of organic matter into a potentially safe, nutrient-stable and sanitary fertilizer. During aerobic composting, temperatures can rise to up to 80°C, a temperature that eliminates most pathogens (Riddech et al. 2002). A number of studies (e.g. Tiquia et al. 1998; Mc Carthy et al. 2011; Gurtler et al. 2018) have reported elimination of pathogens such as Salmonella, Coliforms and faecal Streptococci in contaminated manure after aerobic composting. Moreover, it was noted that antibiotics are degraded during composting (Dolliver et al. 2008; Ramaswamy et al. 2010; Ho et al. 2013; Massé et al. 2014), thus averting potential environmental contamination.

Composting is seen not only as an environmentally acceptable method of waste treatment, but also as one of the more efficient methods of waste disposal that enables recycling of organic matter (FAO 2003). During composting, stable N compounds are formed in the microbial decomposer biomass; these compounds are less susceptible to losses via volatilization, denitrification and leaching. Therefore, the stable N in composted manure has a smaller environmental footprint and is a better source of nutrients when applied to soil as fertilizer for crop production. Furthermore, the partially humified product of composting is relatively resilient to decomposition, therefore compost C is retained in the soil for a longer period (Bernal et al. 2009). Even though during the composting process CO₂ is emitted, the global warming potential of CO₂ from manure is considered to be carbon neutral because it originates from CO₂ that had been fixed by plants during photosynthesis and is not of fossil origin (IPCC 2006, 2019b). Finally, if composted manure is used instead of mineral fertilizer, it will reduce the environmental footprint of pig farming by avoiding GHG emissions associated with fertilizer production and use (Paul et al. 2001).

Biogas production

Biogas is a gas mixture consisting primarily of CO₂ and CH₄, which is produced via microbial digestion of organic materials in an anaerobic environment. Biogas production with pig manure is a suitable method for the treatment of this organic waste, yielding biogas as a useful byproduct (e.g. as cooking gas). The digested manure can be used as soil conditioner and nutrient-rich liquid fertilizer that is valuable for crop production. In Uganda, the biogas production technology was introduced over 50 years ago and since then there has been a lot of effort from nongovernmental organizations and the government to promote this technology. However, the adoption has been low primarily attributed to lack of capital for the investment (Walekhwa et al. 2009). Even with initiatives to alleviate financial constraints, it has been noted that those who adopted the biogas technology often abandoned it after a short period (Nabuuma and Okure 2006). Lwiza et al. (2017) explored the drivers of discontinuation of biogas use and noted that among others, failure to sustain livestock production, reduced availability of family labor and inability of households to repair biogas digesters after malfunctioning were the main causes for abandoning this technology. However, digestion and fermentation of manure in biogas digesters can help reduce pathogen survival in manure, especially in conjunction

with post-digestion composting of bio-slurry to promote further pathogen die-off, which could make this technique well suited for climate-smart and safe pig manure management in SSA (Avery et al. 2014).

Vermicomposting

Vermicomposting is a process by which organic waste such as pig manure are fed to worms (e.g. earthworms and larvae of the Black Soldier Fly [*Hermetia illucens*]), yielding two products—the worm biomass (30–60% protein content) which can be used as livestock feed especially for pigs, chicken and fish (Oonincx et al. 2015); and a humus-like material rich in nutrients that can be used as organic fertilizer (Semiya et al. 2015). The worms can reduce odorous volatile substances by up to 100% (Gajalakshmi and Abbasi 2004; Oonincx et al. 2015; Miranda et al. 2019). Komakech et al. (2015) assessed global warming and eutrophication potentials of composting and vermicomposting as compared to daily manure dumping on landfills in Kampala, Uganda, and found that the global warming potential of composting (5.1 CO₂-eq ton⁻¹) and vermicomposting (17.8 CO₂-eq ton⁻¹) was significantly lower than that of dumped manure (56.6 kg CO₂-eq ton⁻¹). Vermicomposting also had the lowest eutrophication potential. In Uganda, vermicomposting is taking shape as a potential environment and health friendly waste management method in urban settings. A case study in Kampala tested small-scale vermicomposting of cattle manure with food waste using the earthworm species *Eudrilus eugeniae* and found that vermicomposting successfully eliminated *Salmonella* spp. contamination from the organic waste (Lalander et al. 2015). However, vermicomposting did not completely reduce other pathogens (e.g. *Enterococcus* spp., bacteriophage X174 or *Ascaris suum ova*) (Lalander et al. 2013), which is why an additional sanitation step like ammonia treatment was suggested (Vinnerås 2007). A large-scale company—ProTeen Marula Agribusiness, www.weareproteen.com—recently signed a memorandum of understanding with the Kampala City Council Authority to establish a vermicomposting plant at the main landfill site in Kampala. The company has been carrying out pilot vermicomposting trials on urban organic waste using Black Soldier Fly larvae. Adoption of this technique by smallholder farmers is still low. However, vermicomposting has been reported to be a viable business for smallholder farmers with potential return to investment of over 200 % (Lalander et al. 2015). More research is necessary to identify needs of farmers and consumers and constraints concerning use and consumption of products from livestock fed vermicomposted pig manure.

Conclusion

Little attention has been given to pig manure management in Uganda so far. While a few papers have been published on pig production systems in Uganda, data on current manure management practices were rarely collected and presented. In Uganda, pig manure is mostly regarded as waste, of little importance and does not require management. However, given the rapid growth of Ugandan pig production, climate-smart and sanitary solutions for pig manure management are urgently needed to ensure sustainable and safe development of the sector. If managed correctly, pig manure has great potential to improve soil fertility. Proper management can also help mitigate GHG emissions, provide additional income and reduce health risks related to the spread of disease or development of antibiotic resistance. Composting, biogas and vermicomposting are promising manure management solutions for both smallholder and industrial pig producers in Uganda.

References

- Alegbeleye, O.O., Singleton, I. and Sant'Ana, A.S. 2018. Sources and contamination routes of microbial pathogens to fresh produce during field cultivation: a review. *Food Microbiology* 73:177–208.
- Alvåsen, K. 2007. *Farmers' perceptions and handling of livestock manure in urban/peri-urban areas of Kampala, Uganda*. SLU: Department of Animal Nutrition and Management.
- Anderson, J., Learch, C. and Gardner, S. 2016. *National survey and segmentation of smallholder households in Uganda: understanding their demand for financial, agricultural, and digital solutions*. Working paper. Washington, D.C., USA: Consultative Group to Assist the Poor (CGAP).
- Aneja, V.P., Blunden, J., James, K., Schlesinger, W.H., Knighton, R. et al. 2008. Ammonia Assessment from Agriculture : U.S. Status and Needs. *Journal of Environmental Quality* 37: 515–520. doi:10.2134/jeq2007.0002in
- Avery, L.M., Kenneth, Y.A., Vianney, T., Norval, S. and Peter, J.G. 2014. Potential for Pathogen reduction in anaerobic digestion and biogas generation in Sub-Saharan Africa. *Biomass and Bioenergy* 70(1):112–124. doi: 10.1016/j.biombioe.2014.01.053 0961-9534/a
- Bernal, M.P., Alburquerque, J.A. and Moral, R. 2009. Composting of animal manures and chemical criteria for compost maturity assessment. A review. *Bioresource Technology* 100: 5444–5453.
- Bicudo, J.R. and Goyal, S.M. 2003. Pathogens and manure management systems: a review. *Environmental Technology* 24: 115–130.
- Borhan, M.S., Mukhtar, S., Capareda, S. and Rahman, S. 2012. Greenhouse gas emissions from housing and manure management systems at confined livestock operations. In: Rebellon, L.F.M. (ed.), *Waste Management*. Colombia, University of Cauca: 259–296.
- Burns, D.A., Aherne, J., Gay, D.A. and Lehmann, C.M.B. 2016. Acid rain and its environmental effects: Recent scientific advances. *Atmospheric Environment* 146: 1–4.
- Butterbach-Bahl K, Baggs EM, Dannenmann M, et al (2013) Nitrous oxide emissions from soils: how well do we understand the processes and their controls? *Philosophical Transactions of the Royal Society B, Biological Sciences* 368: 20130122–20130122. doi: 10.1098/rstb.2013.0122
- Canh, T.T., Verstegen, M.W.A., Aarnink, A.J.A. and Schrama, J.W. 1997. Influence of dietary factors on nitrogen partitioning and composition of urine and feces of fattening pigs. *Journal of Animal Science* 75: 700–706.
- Caro, D., Davis, S.J., Bastianoni, S. and Caldeira, K. 2014. Global and regional trends in greenhouse gas emissions from livestock. *Climate Change* 126: 203–216.
- Chadwick, D., Sommer, S.G., Thorman, R., Fanguiero, D., Cardenas, L. et al. 2011. Manure management: Implications for greenhouse gas emissions. *Animal Feed Science and Technology* 166–167: 514–531. doi: 10.1016/j.anifeedsci.2011.04.036
- Daniel, T.C., Sharpley, A.N. and Lemunyon, J.L. 1998. Agricultural phosphorus and eutrophication: A symposium overview. *Journal of Environmental Quality* 27: 251–257.
- Dione, M., Ouma, E., Roesel, K., Kungu, J., Lule, P. et al. 2014. Participatory assessment of animal health and husbandry practices in smallholder pig production systems in three high poverty districts in Uganda. *Preventive Veterinary Medicine* 117: 565–576.
- Dolliver, H., Gupta, S. and Noll, S. 2008. Antibiotic degradation during manure composting. *Journal of Environmental Quality* 37: 1245–1253.
- Dourmad, J.Y., Guingand, N., Latimier, P. and Sève, B. 1999. Nitrogen and phosphorus consumption, utilisation and losses in pig production: France. *Livestock Production Science* 58: 199–211.
- Elsgaard, L., Olsen, A.B. and Petersen, S.O. 2016. Temperature response of methane production in liquid manures and co-digestates. *Science of the Total Environment* 539: 78–84.
- Fantozzi, F. and Buratti, C. 2009. Biogas production from different substrates in an experimental Continuously Stirred Tank Reactor anaerobic digester. *Bioresource Technology* 100: 5783–5789.

- FAO (Food and Agriculture Organization of the United Nations). 2012. *Pig Sector Kenya*. In: FAO Animal Production and Health Livestock Country Reviews No. 3. Rome, Italy: FAO.
- FAO. 2003. *On-farm composting methods*. In: Land and Water Discussion Paper 2. Rome, Italy: FAO.
- Gajalakshmi, S. and Abbasi, S.A. 2004. Earthworms and vermicomposting. *Indian Journal of Biotechnology* 3: 486–494.
- Galassi, G., Colombini, S., Malagutti, L., Crovetto, G.M. and Rapetti, L. 2010. Effects of high fibre and low protein diets on performance, digestibility, nitrogen excretion and ammonia emission in the heavy pig. *Animal Feed Science and Technology* 161: 140–148
- Gerber, P.J., Steinfeld, H., Henderson, B., Mottet, A., Opio, C. et al. 2013. *Tackling climate change through livestock – A global assessment of emissions and mitigation opportunities*. Rome, Italy: FAO.
- Goopy, J.P. Onyango, A.A., Dickhoefer, U. and Butterbach-Bahl, K. 2018. A new approach for improving emission factors for enteric methane emissions of cattle in smallholder systems of East Africa – Results for Nyando, Western Kenya. *Agricultural Systems* 161: 72–80. doi: 10.1016/j.agsy.2017.12.004
- Gurtler, J.B., Doyle, M., Erickson, M., Jiang, X., Millner, P. and Sharma, M. 2018. Composting to inactivate foodborne pathogens for crop soil application: a review. *Journal of Food Protection* 81: 1821–1837.
- Guthrie, S., Dunkerley, F., Tabaqchali, H., Harshfield, A., Ioppolo, B. and Manville, C. 2018. *The impact of ammonia emissions from agriculture on biodiversity*. DOI: 10.7249/RR2695
- Herrero, M., Henderson, B., Havlík, P., Thornton, P.K., Conant, R.T. et al. 2016. Greenhouse gas mitigation potentials in the livestock sector. *Nature Climate Change* 6: 452–461. doi: 10.1038/nclimate2925
- Ho Y.B., Zakaria, M.P., Latif, P.A. and Saari, N. 2013. Degradation of veterinary antibiotics and hormone during broiler manure composting. *Bioresource Technology* 131: 476–484.
- Ikwap, K., Jacobson, M., Lundeheim, N., Owiny, D., Nasinyama, G. et al. 2014. Characterization of pig production in Gulu and Soroti districts in northern and eastern Uganda. *Livestock Research for Rural Development* 26(4).
- IPCC. 2006. *2006 IPCC Guidelines for National Greenhouse Gas Inventories*. Geneva, Switzerland: IPCC.
- IPCC (Intergovernmental Panel on Climate Change). 2019a. Summary for Policymakers. In: Shukla, P.R., Skea, J., Slade, R., van Diemen, R., Haughey, E. et al. (eds), *Climate Change and Land: an IPCC special report on climate change, desertification, land degradation, sustainable land management, food security, and greenhouse gas fluxes in terrestrial ecosystems*. Geneva, Switzerland: IPCC.
- IPCC. 2019b. Emissions from livestock and manure management. In: *2019 Refinement to the 2006 IPCC Guidelines for National Greenhouse Gas Inventories*. Geneva, Switzerland: IPCC.
- Johansson, M., Emmoth, E., Salomonsson, A. and Albihn, A. 2005. Potential risks when spreading anaerobic digestion residues on grass silage crops—survival of bacteria, moulds and viruses. *Grass and Forage Science* 60: 175–185.
- Jørgensen, H., Prapasongsa, T. and Poulsen, H.D. 2013. Models to quantify excretion of dry matter, nitrogen, phosphorus and carbon in growing pigs fed regional diets. *Journal of Animal Science and Biotechnology* 4:42.
- Kang, Y., Li, Q., Yin, Z., Shen, M., Zhao, H. et al. 2018. High diversity and abundance of cultivable tetracycline-resistant bacteria in soil following pig manure application. *Scientific Reports* 8: 1–13.
- Kiggundu, N., Ddungu, S.P., Wanyama, J., Cherotich, S., Mpairwe, D. et al. 2019. Greenhouse gas emissions from Uganda's cattle corridor farming systems. *Agricultural Systems* 176:102649. doi: 10.1016/j.agsy.2019.102649
- Kimbi, E., Lekule, F., Mlangwa, J., Mejer, H. and Thamsborg, S. 2015. Smallholder pigs production systems in Tanzania. *Journal of Agricultural Science and Technology A* 5 (2015): 47–60.
- Komakech, A.J., Sundberg, C., Jönsson, H. and Vinnerås, B. 2015. Life cycle assessment of biodegradable waste treatment systems for sub-Saharan African cities. *Resources, Conservation & Recycling* 99: 100–110. doi: 10.1016/j.resconrec.2015.03.006
- Kruse, M. and Bell, J.N.B. 1987. Ammonia emissions and their role in acid deposition. *Atmospheric Environment* 21: 1939–1946.
- Lalander, C., Diener, S., Magri, M.E., Zurbrügg, C., Lindström, A. and Vinnerås, B. 2013. Faecal sludge management with the larvae of the black soldier fly (*Hermetia illucens*) - From a hygiene aspect. *Science of the Total Environment* 458–460: 312–318. doi: 10.1016/j.scitotenv.2013.04.033

- Lalander, C.H., Komakech, A.J. and Vinnerås, B. 2015. Vermicomposting as manure management strategy for urban small-holder animal farms - Kampala case study. *Waste Management* 39: 96–103. doi: 10.1016/j.wasman.2015.02.009
- Li, Y., Jing, Y., Zhang, Z., Jiang, D., Zhang, Q., Hu, J. et al. 2020. Kinetics of Methane Production from the Co-Digestion of Cow Dung, Pig Manure and Corn Straw. *Journal of Biobased Materials and Bioenergy* 14(1): 91–97.
- Liu, J., Yan, H., Cao, S.C., Liu, J. and Zhang, H.F. 2018. Effect of feed intake level on the determination of apparent and standardized total tract digestibility of phosphorus for growing pigs. *Animal Feed Science and Technology* 246: 137–143.
- Lupindu AM, Ngowi HA, Dalsgaard A, et al (2012) Current manure management practices and hygiene aspects of urban and peri-urban livestock farming in Tanzania. *Livest Res Rural Dev* 24:
- Lwiza, F., Mugisha, J., Walekhwa, P.N., Smith, J. and Balana, B. 2017. Dis-adoption of household biogas technologies in Central Uganda. *Energy for Sustainable Development* 37: 124–132.
- MAAIF (Ministry of Agriculture, Animal Industry and Fisheries). 2008. *The national livestock census report 2008*. Kampala, Uganda: MAAIF.
- MacLeod, M., Gerber, P., Mottet, A., Tempio, G., Falcucci, A. et al. 2013. *Greenhouse gas emissions from pig and chicken supply chains – A global life cycle assessment*. Rome, Italy: Food and Agriculture Organization of the United Nations (FAO).
- Massé, D.I., Saady, N.M.C. and Gilbert, Y. 2014. Potential of biological processes to eliminate antibiotics in livestock manure: an overview. *Animals* 4: 146–163.
- Mc Carthy, G., Lawlor, P.G., Coffey, L., Nolan, T., Gutierrez, M. and Gardiner, G.E. 2011. An assessment of pathogen removal during composting of the separated solid fraction of pig manure. *Bioresource Technology* 102: 9059–9067.
- Miranda, C.D., Cammack, J.A. and Tomberlin, J.K. 2019. Life-History traits of the black soldier fly, *Hermetia illucens* (L.) (Diptera: Stratiomyidae), reared on three manure types. *Animals* 9:281.
- Moset, V., Wahid, R., Ward, A. and Møller, H.B. 2019. Modelling methane emission mitigation by anaerobic digestion: effect of storage conditions and co-digestion. *Environmental Technology* 40: 2633–2642.
- Nabiky, J. and Kugonza, D.R. 2016. Profitability analysis of selected piggery businesses in peri-urban communities of Kampala, Uganda. *Livestock Research for Rural Development* 28: 1–10.
- Nabuuma, B. and Okure, M.A.E. 2006. Field-Based Assessment of Biogas, Technology: The Case of Uganda. In: Mwakali, J.A. and Taban-Wani, G. (eds), *Proceedings from the International Conference on Advances in Engineering and Technology*. Elsevier: 481–487.
- Nalubwama, S., Vaarst, M., Kabi, F. and Kiggundu, M. 2014. Animal husbandary practices of smallholder organic farmers in Uganda: Challenges and future prospects. In: *Building Organic Bridges*. 4: 1123–1126.
- Nantima, N., Ocaido, M., Davies, J., Dione, M.M., Okoth, E. et al. 2015. Characterization of smallholder pig production systems in four districts along the Uganda-Kenya border. *Livestock Research for Rural Development* 27.
- Ndambi, O.A., Pelster, D.E., Owino, J., de Buissonjé, F. and Vellinga, T. et al. 2019. Manure Management Practices and Policies in Sub-Saharan Africa: Implications on Manure Quality as a Fertilizer. *Frontiers in Sustainable Food Systems* 3: 1–14. doi: 10.3389/fsufs.2019.00029
- Ndung'u, P.W., Bebe, B.O., Ondiek, J.O., Butterbach-Bahl, K., Merbold, L. and Goopy, J.P. 2018. Improved region-specific emission factors for enteric methane emissions from cattle in smallholder mixed crop: Livestock systems of Nandi County, Kenya. *Animal Production Science* 59: 1136–1146. doi: 10.1071/AN17809
- Ndyomugenyi, E.K. and Kyasimire, J. 2015. Pig production in Kichwamba Sub-county, Rubirizi district, Uganda. *Livestock Research for Rural Development* 27.
- Nguyen, Q.H., Le, P.D., Chim, C., LE, N.D. and Fievez, V. 2019. Potential to mitigate ammonia emission from slurry by increasing dietary fermentable fiber through inclusion of tropical byproducts in practical diets for growing pigs. *Asian-Australasian Journal of Animal Sciences* 32(4): 574–584.
- Okello, E., Collins, A. and De Greve, H. 2015. Analysis of performance, management practices and challenges to intensive pig farming in peri-urban Kampala, Uganda. *International Journal of Livestock Production* 6: 1–7.
- Okoli, C.G., Edo, F.A., Ogbuewu, I., Nwajiobi, I.J., Enemor, V.H.A. and Charles, O.I. 2019. Biochemical Values of Pig Dung Collected from Different Locations in Imo State, Southeastern Nigeria. *Asian Journal of Biological Sciences* 12(3): 470–476.

- Oonincx, D.G.A.B., van Huis, A., van Loon, J.J.A. 2015. Nutrient utilisation by black soldier flies fed with chicken, pig, or cow manure. *Journal of Insects as Food and Feed* 1: 131–139. doi: 10.3920/jiff2014.0023
- Paul, J.W., Wagner-Riddle, C., Thompson, A., Fleming, R. and MacAlpine, A. 2001. Composting as a strategy to reduce greenhouse gas emissions. *Climate Change* 2: 3–5.
- Pelster, D.E., Gisore, B., Goopy, J., Korir, D., Koske, J.K. et al. 2016. Methane and Nitrous Oxide Emissions from Cattle Excreta on an East African Grassland. *Journal of Environmental Quality* 45: 1531. doi: 10.2134/jeq2016.02.0050
- Petersen, S.O., Blanchard, M., Chadwick, D., Del Prado, A., Edouard, N. et al. et al. 2013. Manure management for greenhouse gas mitigation. *Animal* 7: 266–282. doi:10.1017/S1751731113000736
- Pfeifer, C., Morris, J., Githoro, E. and Asiimwe, G. 2015. *Pathways for sustainable pig value chain development in Uganda*. Report of a participatory ex-ante environmental assessment workshop, Hoima, Uganda, 30 September 2015. Nairobi, Kenya: ILRI.
- Portejoie, S., Dourmad, J.Y., Martinez, J. and Lebreton, Y. 2004. Effect of lowering dietary crude protein on nitrogen excretion, manure composition and ammonia emission from fattening pigs. *Livestock Production Science* 91: 45–55.
- Prapasongsa, T., Poulsen, H.D. and Jørgensen, H. 2009. Prediction of manure nitrogen and carbon output from grower-finisher pigs. *Animal Feed Science and Technology* 151: 97–110.
- Ramaswamy, J., Prasher, S.O., Patel, R.M., Hussain, S.A. and Barrington, S.F. 2010. The effect of composting on the degradation of a veterinary pharmaceutical. *Bioresource Technology* 101: 2294–2299.
- Riddech, N., Klammer, S. and Insam, H. 2002. Characterisation of microbial communities during composting of organic wastes. In: Insam, H., Riddech, N. and Klammer, S. (eds), *Microbiology of composting*. Springer: 43–51.
- Robinson, T.P., Thornton P.K., Franceschini, G., Kruska, R.L., Chiozza, F. et al. 2011. *Global livestock production systems*. Rome, Italy: FAO and Nairobi, Kenya: ILRI
- Roesel, K., Dohoo, I., Baumann, M., Dione, M., Grace, D. and Clausen, P.H. 2017. Prevalence and risk factors for gastrointestinal parasites in small-scale pig enterprises in Central and Eastern Uganda. *Parasitology Research* 116: 335–345.
- Sakadevan, K. and Nguyen, M.L. 2017. Livestock production and its impact on nutrient pollution and greenhouse gas emissions. In: Sparks, D. (ed), *Advances in agronomy*. Elsevier: 147–184.
- Selle, P.H., Cowieson, A.J. and Ravindran, V. 2009. Consequences of calcium interactions with phytate and phytase for poultry and pigs. *Livestock Science* 124: 126–141.
- Semiyaga, S., Okure, M.A.E., Niwagaba, C.B., Alex, Y.K. and Kansiime, F. 2015. Decentralized options for faecal sludge management in urban slum areas of Sub-Saharan Africa: A review of technologies, practices and end-uses. *Resources, Conservation and Recycling* 104: 109–119. doi: 10.1016/j.resconrec.2015.09.001
- Serrano-Silva, N., Sarria-Guzmán, Y., Dendooven, L. and Luna-Guido, M. 2014. Methanogenesis and methanotrophy in soil: a review. *Pedosphere* 24: 291–307.
- Sommer, S.G., Petersen, S.O., Sørensen, P., Poulsen, H.D. and Møller, H.B. 2007. Methane and carbon dioxide emissions and nitrogen turnover during liquid manure storage. *Nutrient Cycling in Agroecosystems* 78: 27–36.
- Ström, G., Albiñ, A., Jinnerot, T., Boqvist, S., Djurfeldt, A.A. et al. 2018. Manure management and public health: Sanitary and socio-economic aspects among urban livestock-keepers in Cambodia. *Science of the Total Environment* 621: 193–200.
- Thornton, P.K. 2010. Livestock production : recent trends, future prospects. *Philosophical Transactions of the Royal Society B, Biological Sciences* 365: 2853–2867. doi: 10.1098/rstb.2010.0134
- Thys, S., Mwape, K.E., Lefèvre, P., Dorny, P., Phiri, A.M. et al. 2016. Why pigs are free-roaming: communities' perceptions, knowledge and practices regarding pig management and taeniosis/cysticercosis in a Taenia solium endemic rural area in eastern Zambia. *Veterinary Parasitology* 225:33–42.
- Tiquia, S.M., Tam, N.F.Y. and Hodgkiss, I.J. 1998. Salmonella elimination during composting of spent pig litter. *Bioresource Technology* 63: 193–196.
- Venglovsky, J., Sasakova, N., Gregova, G., Papajova, I., Toth, F. and Szaboova, T. 2018. Devitalisation of pathogens in stored pig slurry and potential risk related to its application to agricultural soil. *Environmental Science and Pollution Research* 25(22): 21412–21419.

- Vinnerås, B. 2007. Comparison of composting, storage and urea treatment for sanitising of faecal matter and manure. *Bioresource Technology* 98: 3317–3321. doi: 10.1016/j.biortech.2006.07.011
- von Holy, A., Lindsay, D. and Beuchat, L.R. 2006. Vectors and conditions for preharvest contamination of fruits and vegetables with pathogens capable of causing enteric diseases. *British Food Journal* 108: 38–53.
- Walekhwa, P.N., Mugisha, J. and Drake, L. 2009. Biogas energy from family-sized digesters in Uganda: Critical factors and policy implications. *Energy Policy* 37: 2754–2762.
- Xie, W., Shen, Q., Zhao, F.J. 2018. Antibiotics and antibiotic resistance from animal manures to soil: a review. *European Journal of Soil Science* 69: 181–195.
- Zhu, Y., Merbold, L., Pelster, D., Diaz-Pines, E., Wanyama, G.N. and Butterbach-Bahl, K. 2018. Effect of Dung Quantity and Quality on Greenhouse Gas Fluxes From Tropical Pastures in Kenya. *Global Biogeochemical Cycles* 32: 1589–1604. doi: 10.1029/2018GB005949



The International Livestock Research Institute (ILRI) works to improve food and nutritional security and reduce poverty in developing countries through research for efficient, safe and sustainable use of livestock. Co-hosted by Kenya and Ethiopia, it has regional or country offices and projects in East, South and Southeast Asia as well as Central, East, Southern and West Africa. ilri.org



CGIAR is a global agricultural research partnership for a food-secure future. Its research is carried out by 15 research centres in collaboration with hundreds of partner organizations. cgiar.org